

1 **M.R. Yagüe<sup>1</sup> & A.D. Bosch-Serra<sup>2</sup>. 2013. Slurry field management and ammonia**  
2 **emissions under Mediterranean conditions. *Short Communication. Soil Use and***  
3 ***Management* 29:397-400. DOI: 10.1111/SUM2012.061**

4

5

6 <sup>1</sup> *Department of Soil & Irrigation (EEAD-CSIC Associated Unit), Agrifood Research*  
7 *and Technology Centre Aragon (CITA), Avda. Montañana 930, E-50059, Zaragoza,*  
8 *Spain.*

9 <sup>2</sup> *Department of Environmental & Soil Sciences, University of Lleida (UdL), Avd.*  
10 *Alcalde Rovira Roure 191, E-25198, Lleida, Spain.*

11

12 **Corresponding author:** mryaguue@aragon.es (MR Yagüe).

13

14 **Running title:** Slurry management and ammonia loss

## Abstract

15  
16 In Spain, farmers are interested in applying pig (*Sus scrofa domesticus*) slurry (PS) to  
17 their fields throughout the year. During the spring and summer months ammonia (NH<sub>3</sub>)  
18 volatilization may be high. We studied the potential range of NH<sub>3</sub> losses under a warm  
19 and a hot period of the year, using available field practices, and two strategies: PS  
20 directly incorporated into the soil, in spring (I-spring); and PS applied by splash-plate,  
21 in summer time (SP-summer), both to bare soil. Measurements were conducted, after PS  
22 application, using the micrometeorological mass-balance integrated horizontal flux  
23 method. The cumulative NH<sub>3</sub>-N volatilization was 35% (I-spring) and 60% (SP-  
24 summer) of total ammonium nitrogen applied, and half of the total NH<sub>3</sub>-N losses  
25 happened by 17h and 8h, respectively, after application. Incorporation strategy was less  
26 effective in avoiding NH<sub>3</sub> losses than is described in the literature. This fact has  
27 important consequences for the implementation of NH<sub>3</sub> mitigation measures in  
28 Mediterranean agricultural systems.

29

30 **Keywords:** ammonia-losses, slurry incorporation-method, ammonia volatilization,  
31 splash-plate, fertilizer strategies.

32

## 33 Introduction

34 Spain is the second largest pig producer in Europe and nearly 50% of Spanish pig  
35 production is concentrated in the Ebro river valley (MARM, 2013).

36 Pig slurry (PS) is mainly applied to maize (long cycle) at sowing (April-May) and to  
37 winter cereals at sowing (October-November) or as a sidedressing (February-March),  
38 leaving a gap of 5-6 months when applications are normally avoided. In irrigated areas,  
39 a second crop after the cereal harvest or the extension of the maize (short cycle) sowing  
40 period can cover this gap.

41 Despite the importance of quantifying European NH<sub>3</sub> emissions, little work done under  
42 Mediterranean conditions has been published on the topic (Génermont & Cellier, 1997;  
43 Sanz *et al.*, 2010) and existing articles do not cover the whole annual period, nor the  
44 range of typical conditions.

45 Our objective was to quantify NH<sub>3</sub> volatilization losses from PS applied according to  
46 two different strategies: i) direct incorporation in spring (I-spring), assumed to  
47 approximate to minimum likely losses, and ii) splash-plate application in summer time  
48 (SP-summer), taken to approximate to maximum likely losses.

49

## 50 **Materials and methods**

51 Two experiments were conducted in a representative area of the Ebro valley (41° 44'N,  
52 0° 49'W, altitude 225 m) on bare ploughed soil (Table 1) before cereal establishment.  
53 The first was established on 16-17 May 2007. An incorporation machine was employed.  
54 Pig slurry incorporation was by a tube divided into three hoses (12 outlets), with a total  
55 application width of 4.80 m. Each outlet was located between two shares, the first one  
56 opened a slot in the soil and the one located at the back buried the applied PS at a depth  
57 of about 0.15-0.20 m. The second experiment was established on 2-3 August 2007  
58 where surface PS spreading was by a tank fitted with a splash-plate; it was spread over  
59 the soil without incorporation. For each strategy (Table 2), three replicates were set up  
60 plus a control.

61 Weather in the experimental period has two limitations, in May with minimum average  
62 temperature (T) and maximum average relative humidity (RH), and in August with  
63 maximum T and minimum RH (Fig. 1).

64 The micrometeorological mass-balance integrated horizontal flux method was used,  
65 following the description and procedure given by Wood *et al.* (2000). Each rotating

66 mast supported three passive NH<sub>3</sub> flux samplers mounted at three heights (0.375, 0.75,  
67 and 1.50 m) with the greatest height being 10% of fetch length, in agreement with  
68 estimates by Itier & Perrier (1976), which were confirmed by a previous field test. The  
69 ammonium solution obtained in passive samples was analyzed with a continuous flow  
70 analyser (AA3–Bran+Luebbe). Sampling started immediately after application and  
71 periodicity for the first day was approximately from 1 to 2h, at 3h, from 5 to 8h, at 12h  
72 and before 24h after application. Later on, the intensity of sampling decreased with time  
73 according to the declining of the intensity of NH<sub>3</sub> flux losses (Fig. 2).  
74 Differences between strategies in NH<sub>3</sub>-N cumulative emissions and also as a percentage  
75 of applied NH<sub>4</sub><sup>+</sup>-N, were analysed using analysis of variance and LSMEANS test (p=  
76 0.05). The statistical package SAS V8.2 was used for all statistical analysis.

77

## 78 **Results and discussion**

79 The average for cumulative NH<sub>3</sub> emissions, measured during the experiments, was 35%  
80 of total ammonium nitrogen (TAN) applied (99 kg NH<sub>3</sub>-N/ha, 25% of total N) in the I-  
81 spring strategy and 60% of applied TAN (122 kg NH<sub>3</sub>-N/ha, 42% of total N) in the SP-  
82 summer strategy (Table 2). Half of the maximum NH<sub>3</sub>-N loss was estimated to occur by  
83 17h and 8h after application in I-spring and SP-summer strategies, respectively (Fig. 2).  
84 For the SP-summer strategy the emission was in the upper ranges found in other studies  
85 (e.g. Sommer *et al.*, 2003; Rochette *et al.*, 2009) but according to Misselbrook *et al.*  
86 (2005), the slurry dry matter content in the SP-summer strategy can explain losses  
87 equivalent to 60% of applied TAN. Incorporation did not reduce volatilization as much  
88 as it was expected to, according to the results from other experiments (Huijsmans *et al.*,  
89 2003), probably because dry soil conditions (Table 1) made it difficult to fully bury the  
90 PS, favouring NH<sub>3</sub> gas diffusion through it.

91 In the context of Mediterranean agricultural systems, the official advice to bury PS  
92 spread over land not later than 24h after application would not be fully effective in  
93 reducing NH<sub>3</sub> emissions, as more than 50% of applied TAN could already be lost during  
94 the first 24 hours after application. In addition, the low moisture content in the soil, in  
95 the hottest months of the year, limits the effectiveness of PS incorporation. New  
96 methods must be investigated and they should be orientated to the reduction of slurry-  
97 air contact (e.g. trail hoses) or/and to slurry infiltration enhancement (e.g. light  
98 irrigation). Nevertheless, their effectiveness will be influenced by soil carbonate content  
99 and this aspect needs further practical evaluation. This research is also necessary in the  
100 framework of models such as ALFAM (Søgaard *et al.*, 2002) that do not fully cover the  
101 special aspects that need to be considered in Mediterranean environments.

102

### 103 **Acknowledgements**

104 The authors thank Miguel Izquierdo, Jesús Gaudó, Juan M. Acín, and Benjamín Subías  
105 for their invaluable field management; Marisa Díez for laboratory assistance; and M.  
106 Rosa Teira for assessment of IHF method use. This study was supported by the National  
107 Institute for Agricultural and Food Scientific Research and Technology of Spain-INIA  
108 (projects RTA2004-114 and RTA2010-126) and by FEDER (European Fund for  
109 Regional Development) and FEADER (European Agricultural Fund for Rural  
110 Development) funds.

111 **References**

- 112 G nermont, S. & Cellier, P. 1997. A mechanistic model for estimating ammonia  
113 volatilization from slurry applied to bare soil. *Agricultural and Forest Meteorology*,  
114 **88**, 145-167.
- 115 Huijsmans, J.F.M., Hol, J.M.G. & Vermeulen, G.D. 2003. Effect of application method,  
116 manure characteristics, weather and field conditions on ammonia volatilization from  
117 manure applied to arable land. *Atmospheric Environment*, **37**, 3669-3680.
- 118 Itier, B. & Perrier, A. 1976. Pr sentation d'une  tude analytique de l'advection. I-  
119 Advection li e aux variations horizontales de concentration et de temp rature.  
120 *Annales Agronomiques*, **27**, 111-140.
- 121 MARM. 2013. Anuario de estadística 2011. Available at:  
122 [http://www.magrama.gob.es/estadistica/pags/anuario/2011/AE\\_2011\\_Completo.pdf](http://www.magrama.gob.es/estadistica/pags/anuario/2011/AE_2011_Completo.pdf)  
123 ; accessed 13/5/2013.
- 124 Misselbrook, T.H., Nicholson, F.A. & Chambers, F.B. 2005. Predicting ammonia losses  
125 following the application of livestock manure to land. *Bioresource Technology*, **96**,  
126 159-168.
- 127 Sanz, A., Misselbrook, T.H., Sanz, M.J. & Vallejo, A. 2010. Use of an inverse  
128 dispersion technique for estimating ammonia emissions from surface-applied slurry.  
129 *Atmospheric Environment*, **44**, 999-1002.
- 130 S gaard, H.T., Sommer, S.G., Hutchings, N.J., Huijsmans, J.F.M., Bussink, D.W. &  
131 Nicholson, F. 2002. Ammonia volatilization from field-applied animal slurry - the  
132 ALFAM model. *Atmospheric Environment*, **36**, 3309-3319.
- 133 Sommer, S.G., G nermont, S., Cellier, P., Hutchings, N.J., Olesen, J.E. & Morvan, T.  
134 2003. Processes controlling ammonia emission from livestock slurry in the field.  
135 *European Journal Agronomy*, **19**, 465-486.

- 136 Rochette, P., Angers, D.A., Chantigny, M.H., MacDonald, J.D., Gasser, M.O. &  
137 Bertrand, N. 2009. Reducing ammonia volatilization in a no till soil by  
138 incorporating urea and pig slurry in shallow bands. *Nutrient Cycling in*  
139 *Agroecosystem*, **84**, 71-80.
- 140 Wood, C.W., Marshall, S.B. & Cabrera, M.L. 2000. Improved method for field-scale  
141 measurement of ammonia volatilization. *Communications in Soil Science and Plant*  
142 *Analysis* **31**, 581–590.

143 **Figure Tables**

144

145 **Table 1.** Physico-chemical soil average characteristics (0-0.30 m)

146

147 **Table 2.** Main characteristics<sup>a</sup> and rates of the pig slurries applied to each plot  
148 according to the two strategies: Incorporation in spring (I-spring) and Splash-plate  
149 in summer (SP-summer). Length of measurement periods and absolute emissions  
150 are included.

151

152

153 **Figure Legends**

154

155 **Figure 1** (a) Wind speed average (m/s); (b) Relative humidity averages of air (%): maximum  
156 (RHmax), medium (RHmed) and minimum (RHmin); (c) Air temperature averages (°C):  
157 maximum (Tmax), medium (Tmed) and minimum (Tmin), on a daily basis for a period  
158 from 2004 to 2010. The black round symbols (●) are associated to average meteorological  
159 data for the period of measurements in each experiment (May and August of 2007). There is  
160 a marked contrast between the two measurement periods and no rainfall occurred in  
161 either of the periods.

162

163 **Figure 2.** Cumulative ammonia emission (NH<sub>3</sub>-N) as a percentage of total ammonium  
164 nitrogen applied (NH<sub>4</sub><sup>+</sup>-N, TAN) and measured in two different strategies: slurry  
165 incorporated in May and slurry spread (splash-plate method) in August. Both trends  
166 were adjusted (\*\*\*, p<0.001) to the Michaelis-Menten equation  
167 [  $N_{\text{NH}_3}(t) = N_{\text{max}} * (t / (t + K_m))$  ], where: N<sub>NH<sub>3</sub></sub> (% of TAN) is the accumulative



168 ammonia loss at time (t);  $N_{\max}$  (% of TAN) is the maximum amount of  $\text{NH}_3\text{-N}$  lost  
169 and  $K_m$  (hours) is the time to reach half of the total losses].

170 **Table 1.** Physico-chemical soil average characteristics (0-0.30 m)

Characteristics	Value
pH (potentiometry 1:2.5) <sup>a</sup>	8.3
Humidity (105° C, % w/w) <sup>b</sup>	1.8
Organic matter (Walkley-Black,% w/w)	2.1
Carbonates (Calcimeter Bernard method, % w/w)	39.0
Sand (0.05-2 mm, % w/w)	25.1
Silt (0.002-0.05 mm, % w/w)	53.6
Clay (<0.002 mm, % w/w)	21.3

171 <sup>a</sup> 1:2.5; 1 soil: 2.5 distilled water (v/v).

172 <sup>b</sup> Soil water content was similar in both periods of slurry application.

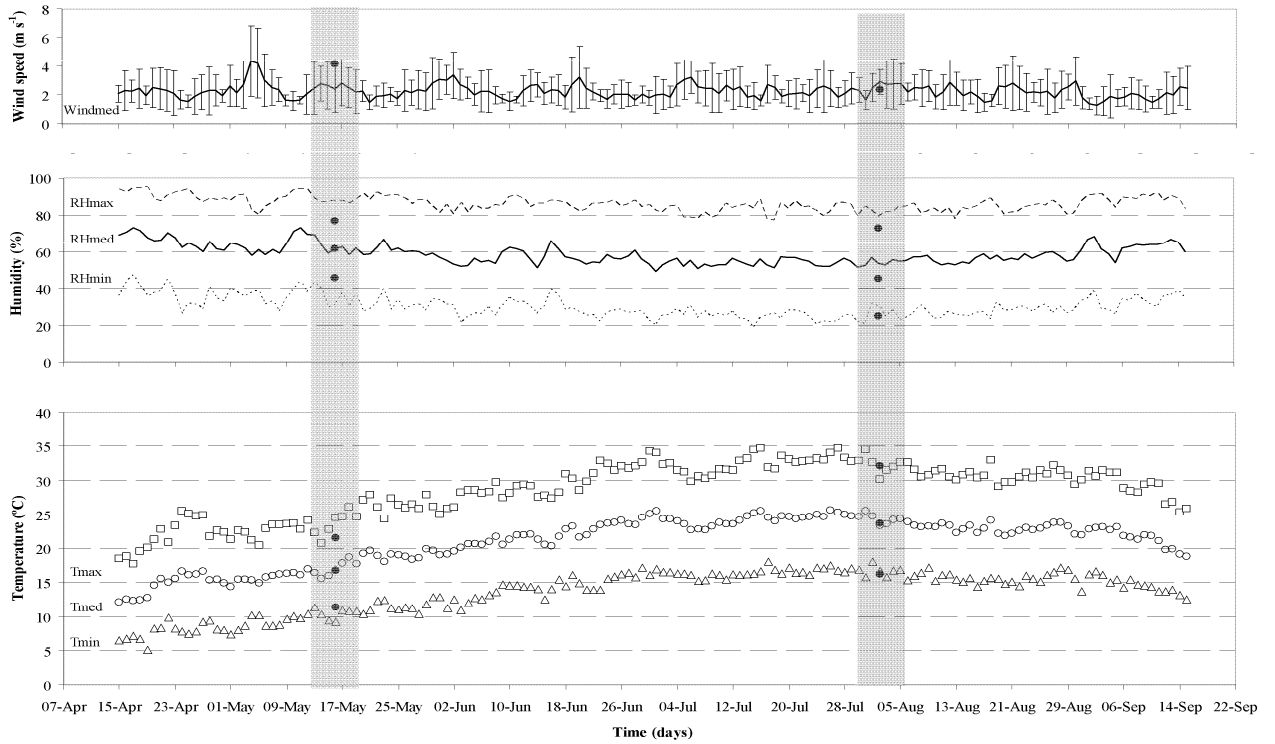
173

174 **Table 2.** Main characteristics<sup>a</sup> and rates of the pig slurries applied to each plot  
 175 according to the two strategies: Incorporation in spring (I-spring) and Splash-plate in  
 176 summer (SP-summer). Length of measurement periods and absolute emissions are  
 177 included.

Strategy	pH (1:5)	Dry matter (kg/t)	Organic matter (kg/t)	Ammoniacal-N (kg NH <sub>4</sub> <sup>+</sup> -N/ha)	Total N (kg N/ha)	Rate (t/ha)	Sampling period (days)	NH <sub>3</sub> -N loss (kg N/ha)
I-Spring	8.2	84.8	59.5	267	377	53	6 days	100.7
I-Spring	8.2	84.8	59.5	302	427	60	7 days	97.6
I-Spring	8.2	84.8	59.5	272	385	54	6 days	99.0
SP-Summer	8.1	86.4	60.2	168	239	36	10 days	102.0
SP-Summer	8.1	86.4	60.2	254	359	54	9 days	157.9
SP-Summer	8.1	86.4	60.2	187	266	40	10 days	106.3

178 <sup>a</sup> pH by potentiometry (1:5; 1 pig slurry: 5 distilled water); dry matter by  
 179 gravimetric analysis at 105°C; organic matter by calcination at 550°C,  
 180 Ammoniacal-N by modified Kjeldahl method and total N by Kjeldahl method.

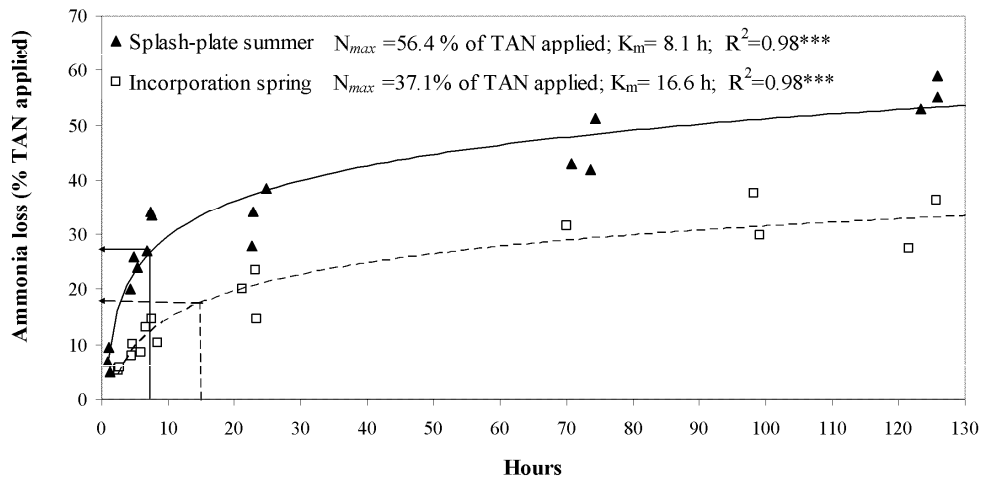
181



182

183 **Fig. 1**

184



185

186 **Fig. 2**