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Putting pyrodiversity to work for animal conservation

Submitted version

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1 **TITLE: Putting pyrodiversity to work for animal conservation**

2 **Introduction**

3 An influential concept in ecology commonly guides conservation decision makers: that
4 environmental heterogeneity drives biodiversity (Stein et al. 2014). In the context of fire
5 management for animal conservation, this concept has encouraged heterogeneity in fire
6 regimes under the assumption that ‘pyrodiversity promotes biodiversity’ (Martin & Sapsis
7 1992).

8 Parr and Andersen’s (2006) important critique of the pyrodiversity hypothesis argued that it
9 was largely unknown how different patterns of burning influenced biodiversity.

10 Subsequently, research on pyrodiversity surged. We recognise that animal conservation in
11 fire-prone ecosystems requires environmental heterogeneity. However, research over the
12 last decade shows that increasing pyrodiversity does not universally promote biodiversity.
13 The misapplication of this assumption, that pyrodiversity promotes biodiversity, can
14 negatively impact animal populations.

15 We reason that applying appropriate levels of pyrodiversity for animal conservation requires
16 (1) recognising that context is important – there is no one-size-fits-all approach, (2)
17 understanding the different mechanisms underpinning the overarching pyrodiversity
18 hypothesis, (3) focusing on ‘functional landscape heterogeneity’ where management of fires
19 is based on species’ demonstrated habitat requirements, and (4) using robust decision
20 analyses that link measurable conservation objectives with strategies to achieve them.

21 **What is pyrodiversity?**

22 Fire regimes vary in the intervals between fires, the seasons when fires occur, the spatial
23 arrangement of fires and their type, severity and intensity. Variation in fire regimes has
24 been termed *pyrodiversity* (Martin & Sapsis 1992). Recent work emphasises how feedbacks
25 between fires, biodiversity and ecosystems influence this variation (Bowman et al. 2016).
26 Because animal species may depend on resources that vary spatially and temporally in
27 response to fires, it is argued that heterogeneous fire regimes provide a range of resources
28 that enable the persistence of a diverse community (reviewed by Parr & Andersen 2006).
29 The expectation of the *pyrodiversity hypothesis* is that biodiversity will increase as
30 spatiotemporal variation in fire increases (Martin & Sapsis 1992). This hypothesis has also
31 been described using the terms ‘fire mosaic’, ‘patch mosaic burn’, ‘successional mosaic’,
32 ‘shifting mosaic’ and ‘vegetation mosaic’ (Supporting Information).

33 **What’s wrong with the general ‘pyrodiversity hypothesis’?**

34 Several studies demonstrate that increasing pyrodiversity does not necessarily increase
35 biodiversity. For example, research in semi-arid woodlands challenged the unqualified
36 application of ‘pyrodiversity promotes biodiversity’ because some fire age-classes provide
37 disproportionately important habitat for vertebrates (Taylor et al. 2011; Kelly et al. 2012;
38 Nimmo et al. 2013). Work in tropical savannas shows that the diversity of termites in South
39 Africa (Davies et al. 2012) and ants in Australia (Andersen et al. 2014) are resilient to
40 changing levels of pyrodiversity – strengthening the thesis of Parr and Andersen (2006) that
41 not all patterns of fires are ecologically meaningful.

42 Other studies show biodiversity can increase with pyrodiversity. For example, pollinator
43 diversity increased with pyrodiversity in mixed-conifer forest in North America (Ponisio et al.
44 2016). The number of fire age-classes increased bird diversity in temperate Australian

45 forests (Sitters et al. 2014). Spatiotemporal variation in fire and grazing regimes enhanced
46 populations of birds, insects and mammals in grasslands in North America (Fuhlendorf et al.
47 2009). Finally, ant diversity in tropical savanna in South America was promoted by diversity
48 in the frequency and season of fires (Maravalhas & Vasconcelos 2014).

49 This snapshot of studies shows that the relationship between pyrodiversity and biodiversity
50 depends on context, being influenced by taxa, ecosystem, scale and location (more
51 examples in Supporting Information). How biodiversity is measured is also important
52 (Giljohann et al. 2015). Given the inadequacy of the general pyrodiversity model, how can
53 scientists and decision makers achieve desirable outcomes for animal conservation?

54 **A pragmatic approach for achieving desirable levels of pyrodiversity**

55 *Recognise the importance of context*

56 Natural ecosystems are not uniform. They contain different species, have different fire
57 regimes and fuels, and present different fire risks to people and biodiversity (DellaSala &
58 Hanson 2015). General statements that pyrodiversity promotes biodiversity ignore this
59 critical detail. A desirable level of pyrodiversity in one context is quite different to a
60 desirable level in another. Martin and Sapsis (1992) noted carefully that ‘It makes no sense
61 to have the same policy for all [vegetation] types, regardless of the role of fires in them’.
62 Scientists should not gloss over this important detail when communicating with each other,
63 decision makers and the public.

64 *Differentiate between different hypotheses*

65 The pyrodiversity hypothesis is not just one hypothesis – it is many. It is based on several
66 mechanisms concerning how variation in fire regimes enhances the persistence of individual
67 species and the coexistence of multiple species. This includes variation in fires related to
68 compositional heterogeneity, configurational heterogeneity, temporal change, and
69 interactions between fires and other processes (Figure 1; Supporting Information). An
70 overarching pyrodiversity hypothesis has usefully galvanised fire research but it has also led
71 to difficulties when comparing studies done in different places and at different scales.
72 Multiple underlying mechanisms make it difficult to communicate particular effects under
73 the general pyrodiversity hypothesis. This in turn makes it difficult to implement effective
74 management for animal conservation.

75 One way to progress pyrodiversity research is to better clarify the sub-hypotheses or
76 alternative hypotheses that are tested. Recent pyrodiversity studies have tested
77 mechanisms related to *habitat amount*, *habitat complementation*, *habitat fragmentation*,
78 *habitat heterogeneity*, *habitat refuges* and variation in *fire season* (Supporting Information).
79 Such mechanisms and hypotheses must be organised more clearly to guide management.
80 More useful outputs for animal conservation will now be developed by asking specific
81 questions such as “How does spatiotemporal variation in the fire regime influence biota?”
82 than by asking “Does pyrodiversity promote biodiversity?”.

83 *Focus on functional heterogeneity*

84 Measures of fires without reference to particular animal species traditionally underpin fire
85 management (Clarke 2008). Most fire management plans are based on measures that
86 represent ‘structural landscape heterogeneity’ (*sensu* Fahrig et al. 2011) where landscape
87 elements are identified by physical characteristics without reference to particular taxa (e.g.

88 satellite images). But changes in structural heterogeneity, such as the number of fire age
89 classes, do not necessarily represent changes in animal diversity. Consequently, fire
90 management for biodiversity conservation should be based on the demonstrated
91 requirements of animals *and* the plants, habitats and landscape elements they depend on:
92 'functional landscape heterogeneity' (Fahrig et al. 2011). Functional heterogeneity could be
93 defined by classifying and mapping fire elements based on the known requirements of
94 multiple species and their life histories, such as feeding and nesting sites. For example,
95 Bradstock et al. (2005) proposed a conceptual model that explores functional heterogeneity
96 using combinations of permanent and seral (changing) habitat. Models of how fires
97 influence species' distributions can also define desirable levels of functional heterogeneity,
98 and these empirical outputs used to identify fire-sensitive species and to inform ecosystem
99 management (Kelly et al. 2015). In each case, differentiating between sub-hypotheses of the
100 general model helps to define desirable levels of functional landscape heterogeneity for
101 biodiversity. A continuing challenge is to ensure functional heterogeneity represents both
102 vertebrates and invertebrates.

103 *Implement decision frameworks that consider uncertainty*

104 How can desirable levels of pyrodiversity be achieved over time and while considering
105 uncertainty such as unplanned fire? One approach uses decision tools that encourage
106 conservation managers to explicitly state objectives and alternative management options
107 while accounting for uncertainty (Driscoll et al. 2010; McCarthy 2014). For example, recent
108 work demonstrates that if fire managers aim to maximise biodiversity in semi-arid
109 woodlands then they need to make different decisions depending on the state of the
110 landscape. The best prescribed burning strategy for maximizing the relative abundance of

111 22 species of birds, mammals and reptiles depended on how much of the landscape was
112 comprised of early, middle and late successional vegetation (i.e. context) (Giljohann et al.
113 2015).

114 Importantly, fire planning must consider the combined effects of prescribed and unplanned
115 fires. Most fire management plans implicitly assume that unplanned fires will not occur. This
116 myopia to the likelihood of unplanned fires will underestimate the extent of burning, with
117 negative consequences for animal conservation. Decision tools and scenario planning enable
118 the role of unplanned fires to be better understood (Regos et al. 2016). These approaches
119 will be most effective when developed with local stakeholders and traditional land owners.

120 **Challenges for animal conservation**

121 The relationship between pyrodiversity and biodiversity is much better understood
122 following research over the last decade. These advances have helped to define desirable
123 levels of pyrodiversity worldwide (Supporting Information). In that time, fire-prone
124 ecosystems have also experienced significant changes, including more extreme fire weather,
125 growing use of prescribed burning and serious declines in some animal assemblages (Moritz
126 et al. 2014). Now, more than ever, our understanding of animal responses to fire should be
127 used to determine management objectives and actions. We think that these four pragmatic
128 recommendations – emphasising the importance of context, mechanisms, functional
129 heterogeneity and uncertainty - will aid interpretation of pyrodiversity studies and improve
130 conservation efforts globally.

131 **Supporting Information**

132 Alternative terms to describe spatiotemporal variation in fire regimes (Appendix S1) and A
133 'snapshot' of recent pyrodiversity studies (Appendix S2) are available online. The authors
134 are solely responsible for the content and functionality of these materials. Queries (other
135 than absence of the material) should be directed to the corresponding author.

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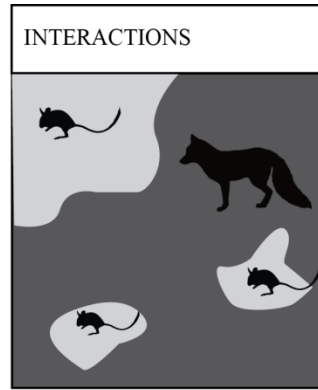
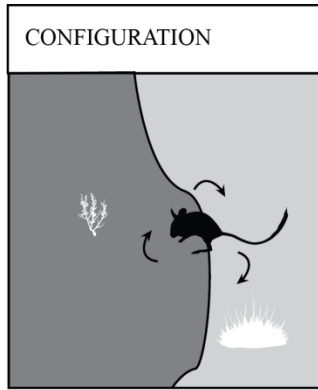
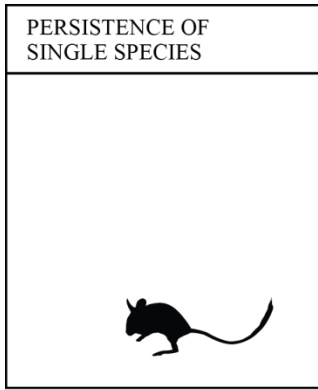
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209 **Figure legend**

210 **Figure 1.** The overarching pyrodiversity hypothesis is underpinned by multiple mechanisms
211 and hypotheses. Mechanisms can be classified according to their influence on the
212 persistence of single species and the coexistence of multiple species. Pyrodiversity can be
213 measured in several ways relating to compositional heterogeneity (the number, amount and
214 type of fire elements), configurational heterogeneity (the spatial arrangement of fire
215 elements), various temporal mechanisms relating to fire interval, duration and season, and
216 interactions with other processes (including climate, disease, fragmentation, grazing and
217 predation). These measures, and their interactions in time and space, have been used to
218 test a range of sub-hypotheses relating to pyrodiversity. Here we highlight four sub-
219 hypotheses that will be important to continue to test, in different contexts, to enhance
220 animal conservation in fire-prone ecosystems: the *habitat complementation hypothesis* –
221 that multiple fire elements within the landscape supplement the requirements of individual
222 species (configurational heterogeneity); the *habitat heterogeneity hypothesis* - that
223 increased variation in fire elements will enhance the coexistence of multiple species
224 (compositional heterogeneity); the *habitat refuge hypothesis* – that the quality and spatial
225 configuration of fire elements influence immediate survival and recolonization after a fire
226 (interactions between fire and predation); and the *fire season hypothesis* - that prescribed
227 burning in cooler months will reduce wildfire size and enhance the survival of animals
228 (temporal variation).

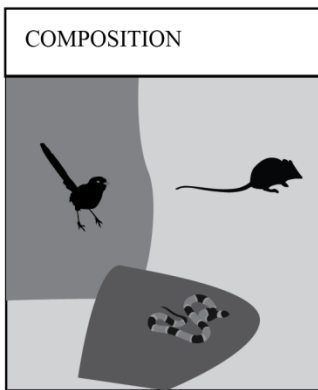
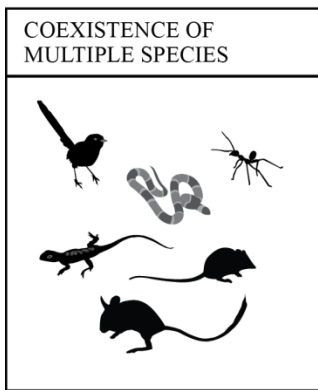
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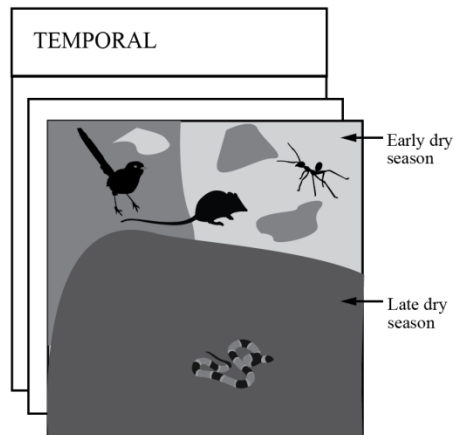


E.g. Habitat complementation hypothesis

E.g. Habitat refuge hypothesis



E.g. Habitat heterogeneity hypothesis



E.g. Fire season hypothesis